

Forest Health Protection



Numbered Report 06-08

July 2006

Evaluation of Establishment of *Cyphocleonus achates* and its Potential Impact on Spotted Knapweed

**Nancy Sturdevant, Entomologist
USDA Forest Service, Forest Health Protection
Northern Region, Missoula, MT**

**Sandy Kegley, Entomologist
USDA Forest Service, Forest Health Protection
Northern Region, Coeur d'Alene, ID**

**Yvette Ortega, Wildlife Biologist
Rocky Mountain Research Station, Missoula, MT**

**Dean Pearson, Research Ecologist
Rocky Mountain Research Station, Missoula, MT**

INTRODUCTION

Environmental characteristics of a site may greatly influence the establishment and impact of biological control agents on the target weed. A biological control agent that is highly successful in one region may be virtually ineffective in another (Gurr and Wratten 2000). To maximize the use of limited availability of agents and financial resources, we should try to improve the success of our biological control releases by understanding the site requirements of individual biological control agents.

Thirteen insects have been released on spotted knapweed in the United States since the early 1970s. Seven insect species appear to be well established and of these, four show promise for biological control of spotted knapweed (Corn et al. 2006). *Cyphocleonus achates* (Fahraeus) was

first introduced into Montana in 1988 by (Story et al. 1994) and is now well established in many counties in the state.

European screening studies showed that *C. achates* attacks spotted and diffuse knapweed (Stinson et al. 1994). The larvae mine the roots of knapweed plants in Montana between about mid-July through late spring. Larval mining causes extensive root damage, reductions in shoot biomass (Steinger & Muller-Scharer 1992) and plant mortality (Corn et al. 2006).

Clark et al. (2001a) found that the probability of establishment of *C. achates* was greatest in continuous knapweed infestations, larger infestations, and at mid-elevational sites as compared to lower and higher elevations.

Sturdevant and Dewey (2003) reported that size of plants, weather conditions during and following the insect release, number of years since the release and number of insects released can influence the probability of establishment of *C. achates*. Recent declines of spotted knapweed in western Montana have been coincidental with the release of *C. achates* and the recent drought. Beginning in 1997, Forest Health Protection in cooperation with other USDA Forest Service and county weed cooperators began to monitor impact of weevils on knapweed at previous release sites.

The objectives of this project were to: 1) evaluate the relationship between environmental variables and the establishment of *C. achates* at release sites to try and identify factors that may determine establishment success, and 2) examine the relationship between *C. achates* presence and abundance and changes in spotted knapweed over time to assess the short-term efficacy of this agent in suppression of its target weed.

METHODS

1997/2002 Data

The primary dataset analyzed was based on 48 study sites sampled in both 1997-98 and 2002-03. Records indicated that *C. achates* had been released at 46 of these sites between 1991 and 1997, with most releases occurring in the 1993-1996 period. An average of 150 weevils was released per site (range 26-445). Two additional sites with no record of *C. achates* release were included in analyses since weevils were detected on these sites during initial sampling. According to records, the biological control agent *Agapeta zoegana* was also released and/or detected at 41 of the study sites within the same timeframe as *C. achates*.

In 1997 and 2002, densities of *C. achates* and *A. zoegana* per study site were assessed by counting

larvae in roots of knapweed plants collected at points equally spaced along four transects radiating from the central release point (n=52/site in 1997-98 and n=20/site in 2002-3). Root sample size was reduced because a separate study showed no significant differences in the average number of larvae when sampling 20 versus 52 roots (Sturdevant and Dewey 2002). Counts of adult *C. achates* from sweep net sampling were used to assess presence of weevils, but otherwise were not considered in analyses since sampling of larvae in roots provides a more consistent measure of abundance (Clark et al. 2001a).

Environmental variables describing site-level characteristics were assessed once during the study period. Variables included: elevation, aspect (reclassified as warm [S, SW, W] versus cold [N, NE, E]), slope, topography (reclassified as valley bottom versus not), disturbance level (reclassified as low [scores 0-1] versus high [scores 2-3]), and primary soil type (reclassified as sandy versus not). Categorical variables were reclassified to reduce the number of classes and increase the number of observations per class, thereby facilitating analyses. However, before doing so, preliminary relationships with *C. achates* prevalence were considered to assure that patterns were not lost as a result of collapsing classes.

Additional environmental variables described vegetation. In 1997 and 2003, cover of bare ground, shrubs, spotted knapweed, and remaining vegetation were visually estimated from the release point within a 1/300-acre circular plot. In addition, density of knapweed stems was measured in both sampling periods within 16 1-m plots radiating out from the release point in the four cardinal directions. In 2002 & 2003 only, average height of mature knapweed plants, density of mature knapweed plants, and cover of knapweed seedlings were also recorded within 16, 1-m plots.

2002-2004 Data

Additional data relating *C. achates* prevalence to knapweed characteristics, 2002-2004, were obtained from a report to FHP by Sturdevant and Dewey (2003). Study sites (n=15) had been chosen from a set of *C. achates* release sites (first monitored in 2002) because they had moderate to high numbers of weevils. At each of these sites, relative abundance of adult *C. achates* was estimated by sweeping a total of 300 meters of knapweed per site. Sweeping began near the release point and continued through the densest portions of knapweed surrounding the release point. Knapweed characteristics including height of mature plants, density of stems, and density of mature plants were measured in 16 1-m plots that radiated out from the release point in the four cardinal directions. All sampling was done approximately during the first 2 weeks in August.

Statistical Analyses

We used the 1997/2002 dataset to relate the degree of *C. achates* establishment to environmental site variables. The degree of *C. achates* establishment per release site was defined and independently analyzed in two ways: 1) presence of *C. achates* as determined by detection of *C. achates* larvae or adults in 2002, and 2) density of *C. achates* larvae in 2002. We used logistic regression to evaluate the presence of weevils as a function of background environmental variables (i.e., excluding knapweed characteristics). Similarly, we evaluated the relationship between density of weevils and background environmental variables using mixed linear models.

In a second analysis, we examined knapweed characteristics measured in 1997 and 2002, testing for: 1) association with *C. achates* establishment in terms of presence of weevils and density of larvae, respectively, 2) population trends (i.e., differences between years), and 3) influence of *C. achates* establishment on population trends (i.e., interaction of *C. achates*

variables with year effects). Mixed linear models were used to conduct tests, with year treated as a repeated measure.

Knapweed characteristics measured from 2002-2004 were analyzed with the same mixed model used in the 1997/2002 analysis, i.e., testing knapweed variables for association with *C. achates*, for population trends, and for influence of *C. achates* on population trends. Total number of adult weevils caught in sweep nets was used to represent the prevalence of *C. achates* per site and year since larval counts were not available.

All statistical analyses were conducted with SAS (SAS Institute 1999). We report least squares means from mixed models instead of raw means to adjust for imbalances in sample size and account for influence of covariates.

RESULTS

1997/2002 Data

Densities of *C. achates* increased substantially at release sites between 1997-98 and 2002-03, from an average of $0.08 \pm \text{SE}$ of 0.067 larvae/root to 0.3 ± 0.068 larvae/root ($P=0.012$). Neither release date, number of *C. achates* released, nor number of releases were significantly related to densities of *C. achates* larvae in 2002 ($P>0.4$). However, densities of *C. achates* were positively associated with densities of *A. zoegana* ($P=0.01$).

C. achates was detected at 36 of 48 release sites in 2002. Multivariate analysis of background environmental variables indicated that primary soil type and topography were associated with the presence of *C. achates* in 2002: weevils tended to occur at sites with sandy soils ($P=0.026$) located in valley bottoms ($P=0.046$). Remaining background environmental variables showed no significant relationships with the presence of weevils ($P>0.37$). Sites with sandy soils tended to have higher densities of *C. achates* larvae in 2002, but the difference was only marginally significant in a multivariate test

($P=0.061$) and less so in a bivariate test ($P=0.14$). Remaining background environmental variables showed no association with the density of *C. achates* larvae ($P>0.16$).

There was a trend of less knapweed cover at sites where *C. achates* was present but it was not significantly different ($P=0.57$) and did not vary significantly with the density of weevil larvae ($P=0.49$). Knapweed cover declined significantly between years, from an average of

$49.7 \pm 3.94\%$ in 1997 to $41.7 \pm 4.09\%$ in 2002 ($P=0.02$). The decline in knapweed cover over this time period tended to be greater at sites where *C. achates* was present compared to sites where it was absent, but this difference was not statistically significant ($P=0.16$; Fig. 1). When soil type and topography (i.e., variables that were significantly associated with the presence of *C. achates*) were included in statistical models, results were not altered.

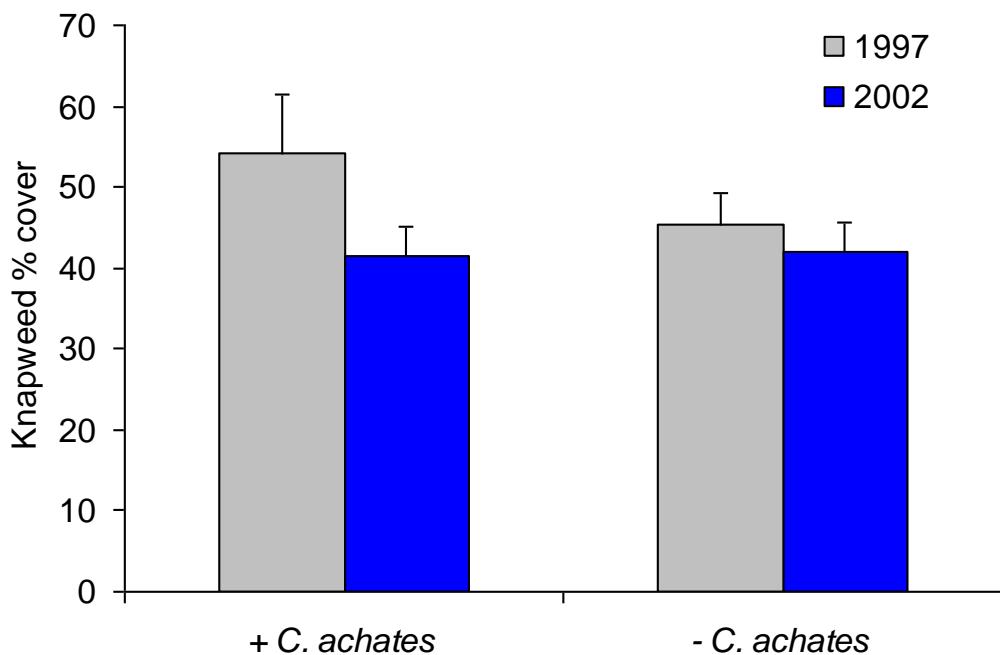


Figure 1. Declines in mean percent cover of spotted knapweed from 1997 to 2002 at release sites where *C. achates* was present (+ *C. achates*) versus absent (- *C. achates*) as determined by detection of weevils during 2002 sampling.

Density of mature spotted knapweed stems across years did not differ significantly between sites where *C. achates* was present versus absent ($P=0.47$) and did not vary significantly with the density of weevil larvae ($P=0.72$). Stem density declined between 1997 and 2002, from an average of 31.4 ± 3.42 stems/m² to 11.1 ± 2.51

stems/m² ($P<0.001$). The decline in stem density from 1997 to 2002 was not affected by the presence of weevils ($P=0.44$; Fig. 2) or by larval densities ($P=0.63$). Inclusion of soil type and topography in statistical models did not affect results.

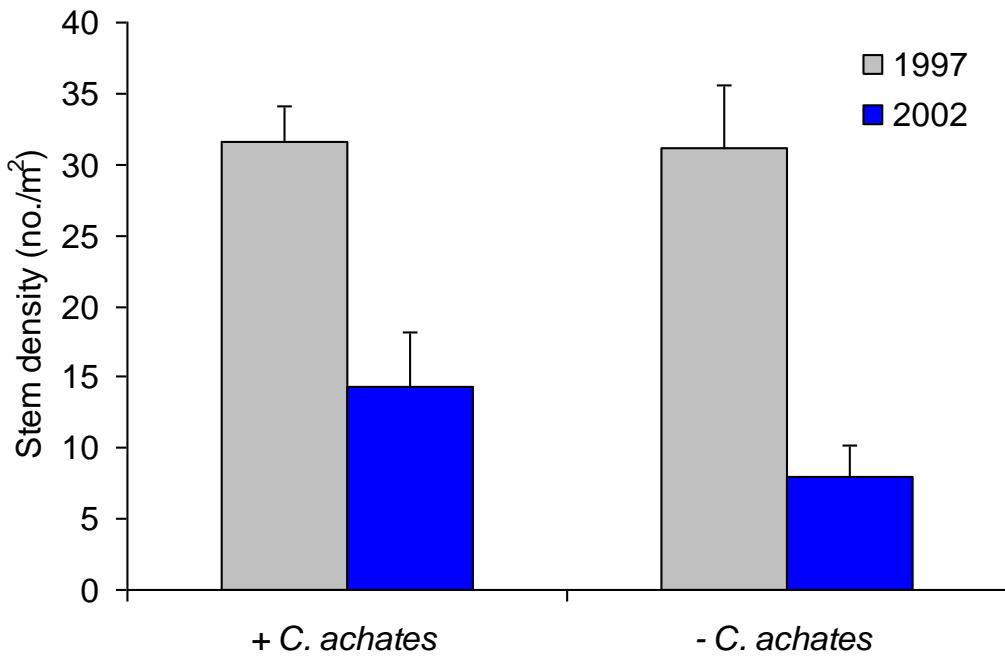


Figure 2. Declines in mean stem density of spotted knapweed from 1997 to 2002 at release sites where *C. achates* was present (+ *C. achates*) versus absent (- *C. achates*) as determined by detection of weevils during 2002 sampling.

2002-2004 Data

Average number of adult *C. achates* detected per site was 11.6 ± 5.38 weevils in 2002, 21.4 ± 5.38 weevils in 2003, and 6.9 ± 5.53 weevils in 2004, but differences among years were only marginally significant ($P=0.08$). For each site, counts of weevils conducted in 2002 and 2003 were highly correlated ($r=0.7$, $P=0.004$). However, weevil counts from 2004 showed no relationship with prior years ($r<0.3$, $P>0.3$), likely reflecting the fact that sampling conditions in 2004 were particularly unfavorable, as noted by Sturdevant and Dewey (2003). Given results of correlation analysis and the tendency for higher counts in 2003, we used the 2003 sample to characterize the relative abundance of *C. achates* at study sites in remaining analyses.

Height of spotted knapweed plants across years did not vary significantly with the relative abundance of *C. achates* at sites ($P=0.62$).

Although plant height had a tendency to decline between 2002 and 2004, the trend was non-significant ($P=0.88$). Furthermore, the yearly variation in plant height was not affected by the relative abundance of *C. achates* ($P=0.59$). Neither stem density nor plant density across years showed a significant relationship with the relative abundance of weevils at sites ($P>0.4$). Both stem and plant density declined significantly between 2002 and 2004 ($P<0.03$), but the degree of decline was not significantly influenced by the relative abundance of *C. achates* ($P>0.5$).

DISCUSSION

This study showed that *Cyphocleonus achates* can establish across a wide variety of sites if given time. *C. achates* also appears to have generally increased in numbers at most release sites since its introduction. In the 1997/2002 dataset, *C. achates* increased by nearly four-fold over the study period, as measured by larval densities within knapweed roots. There was also

some evidence of an increase in *C. achates* in the 2002-2004 dataset, as measured by sweep-netting adult weevils. Although weevil counts were low in 2004, this was likely due to the prevalence of cool, rainy conditions during the sampling period which likely caused *C. achates* to be inactive (Sturdevant and Dewey 2003). In contrast, sampling during more normal conditions in 2002 and 2003 indicated that weevils had roughly doubled between years.

Our observations and analysis of the 1997/2002 dataset indicated that establishment of *C. achates* tended to be associated with sandy soils and valley bottom sites, but not with elevation, aspect, slope, disturbance level, or cover of bare ground, shrubs, or other non-knapweed vegetation. In their study of the same general set of monitoring sites that we examined, Clark et al. (2001b) reported that *C. achates* establishment was most commonly associated with mid-elevations and found no relationship with soil conditions. However, their analyses were based on data collected in 1997-1998 only, i.e., early in the study period. In the current study, *C. achates* was detected at only 19 of the 48 (40%) release sites in 1997 compared to 36 of the 48 (75%) release sites in 2002. Also, 4 of the 19 sites with evidence of establishment in 1997 showed no evidence of weevils in 2002. Given this substantial difference in the subset of sites where *C. achates* appeared to have established in 1997 versus 2002, it is not surprising that Clark et al.'s (2001b) results differed from ours.

Several measures of spotted knapweed viability suggested that knapweed declined from 1997 to 2002 and from 2002 to 2004. Percent cover of knapweed and knapweed stem density showed significant reductions from 1997 to 2002, and stem densities and plant densities declined significantly from 2002 to 2004. Knapweed cover from 1997 to 2002 appeared to decline more markedly at sites where *C. achates* was present (Fig. 1), but this difference was not statistically significant. Knapweed cover was assessed at a relatively broad spatial scale (1/300 acre), which may have resulted in less precise

estimates, making resolution of patterns more difficult. However, in contrast to apparent trends in knapweed cover, declines in stem density over the study period were strong whether or not *C. achates* was present at study sites. Corn et al. (2006) found no difference between number of stems and seedheads between treatments with high versus low numbers of weevils. Other studies have shown that number of stems is not a good indicator of knapweed vigor. Finally, we found no significant relationships between measures of *C. achates* abundance and knapweed declines in either dataset.

We also examined the potential effects of *Agapeta zoegana*, another agent for which data were collected, on knapweed decline. The analysis showed no significant association with knapweed decline. The other most prominent biological control agents at the sites would be the ubiquitous gall flies (Story et al. 1987). However, several intensive studies have indicated that these agents do not reduce spotted knapweed populations (e.g., Maddox 1982, Stanley 2005). The seedhead weevils, *Larinus* species, would be the only other biological control agents likely to occur in any numbers at multiple sites, but even now they are not well distributed. For example, *Larinus* was only present at 23% of the *C. achates* release sites in the 1997/2002 dataset.

From these data, it appears that spotted knapweed has declined over the monitoring period and that a combination of factors may have influenced this decline. Numerous authors across the region have recently reported declines in spotted knapweed abundance corresponding with spring drought conditions occurring in the early 2000s (Ortega et al. 2004, Stanley 2005, Pearson and Callaway 2006, and historical studies have also reported declines in spotted knapweed in association with past droughts occurring prior to the introduction and establishment of most knapweed biological control agents (e.g., Schirman 1981).

In light of these conclusions, it is important to note the limitations of these monitoring data for evaluating cause and effect relationships. For example, to be able to separate the effects of *C. achates* establishment from other factors affecting knapweed, control sites lacking weevils are needed to quantify background trends in knapweed populations. In our analysis of the 1997/2002 dataset, control sites were limited to those sites where *C. achates* had failed to establish. This resulted in a smaller sample than would be ideal given what appeared to be a confounding effect of other factors such as drought. Furthermore, measures of relative abundance derived from sweep netting (as in the case of the 2002-2004 dataset), were sensitive to weather conditions, and therefore may not have adequately captured trends in *C. achates* density. Future studies should rely on larval monitoring in roots for providing a consistent population measurement.

We found that *C. achates* has increased at monitoring sites since it was introduced and that spotted knapweed has declined during this same period. Trends found in this study showed that there was some evidence that *C. achates* may have impacted spotted knapweed cover, as discussed above. Story et al. (2006) showed that spotted knapweed density declined significantly over an eleven-year period at two sites after *C.*

achates increased dramatically at both sites.

Although the trend in our study was not statistically significant, it is possible that *C. achates* was beginning to weakly impact spotted knapweed, since it was released for a relatively short period of time prior to the start of this study. Also, our study included sites where weevil numbers ranged from low to high. Corn et al. (2006) found that plants may survive initial attacks by large numbers of weevils, but may be stressed and continue to die over time. They also found that knapweed mortality occurred only when weevils were found in high numbers and/or over multiple years of attack. Our study showed a decline in plant density over time in the 2002/2004 data set. All of the sites in this data set had moderate to high numbers of weevils.

In summary, *C. achates* established at most sites (75%) over five years and establishment was highest at sites with sandy soils and located on valley bottoms. Knapweed density decreased across sites, but decreases were not significantly correlated with *C. achates*. However, the effects of the biological control agents may have been overshadowed by the effects of the recent drought which probably also played a role in the decrease in knapweed plant density.

REFERENCES

Clark, S. E., R. G. V. Driesche, N. Sturdevant, J. Elkinton, and J. P. Buonaccorsi. 2001a. Effects of site characteristics and release history on establishment of *Agapeta zoegana* (Lepidoptera: Cochylidae) and *Cyphocleonus achates* (Coleoptera: Curculionidae), root-feeding herbivores of spotted knapweed, *Centaurea maculosa*. *Biological Control* 22:122-130.

Clark, S. E., R. G. V. Driesche, N. Sturdevant, and S. Kegley. 2001b. Comparative efficacy of adult and larval sampling for detection of the root-boring insects *Agapeta zoegana* (Lepidoptera: Cochylidae) and *Cyphocleonus achates* (Coleoptera: curculionidae) released for biological control of spotted knapweed. *Biological Control* 30:589-594.

Corn, J. G., J. M. Story and L. J. White. 2006. Impacts of the biological control agent *Cyphocleonus achates* on spotted knapweed, *Centaurea maculosa*, in experimental plots. *Biological Control*. 37:75-81.

Gurr, G., and Wratten, S. 2000. *Biological Control: Measures of Success*. Kluwer Academic Publishers, Boston, MA.

Maddox, D. M. 1982. Biological control of diffuse knapweed (*Centaurea diffusa*) and spotted knapweed (*Centaurea maculosa*). *Weed Science* 30:76-82.

Ortega, Y. K., D. E. Pearson, and K. S. McKelvey. 2004. Effects of exotic plant invasion and introduced biological control agents on native deer mouse populations. *Ecological Applications* 14:241-253.

Pearson, D. E., and R. M. Callaway. 2006. Biological control agents elevate hantavirus by subsidizing mice. *Ecology Letters*: in press

Pearson, D. E., and R. M. Callaway. 2005. Indirect nontarget effects of host-specific biological control agents: implications for biological control. *Biological Control* 35:288-298.

Pearson, D. E. and R. J. Fletcher, Jr. *In review*. Exotic organisms as food subsidies: removal of biological control agents reduces consumer populations.

SAS Institute. 1999. *SAS/STAT User's Guide, Version 8*. SAS Institute Inc., Cary, North Carolina.

Schirman, R. 1981. Seed production and spring seedling establishment of diffuse and spotted knapweed. *Journal of Range Management* 34:45-47.

Shea, K., D. Kelly, A. W. Sheppard, and T. L. Woodburn. 2005. Context-dependent biological control of an invasive thistle. *Ecology* 86:3174-3181.

Stanley, A.G. 2005. Evaluating the effectiveness of biological control: spotted knapweed, seed head gallflies, predacious mice, and environmental variation. Doct. Diss. University of Washington, Seattle, WA.

Steinger, T., and H. Müller-Schärer. 1992. Physiological and growth responses of *Centaurea maculosa* (Asteraceae) to root herbivory under varying levels of interspecific competition and soil nitrogen availability. *Oecologia* 91:141-149.

Story, J. M., R. M. Nowierski, and K. W. Boggs. 1987. Distribution of *Urophora affinis* and *U. quadrifasciata*, two flies introduced for biological control of spotted knapweed (*Centaurea maculosa*) in Montana. *Weed Science* 35:145-148.

Story, J. M., W. R. Good and L. J. White. 1997. First report of establishment of *Cyphocleonus achates* (Fahr.) (Coleoptera:Curculionidae) on spotted knapweed, *Centaurea maculosa* Lamarck, in the United States. The Canadian Entomologist. 129:373-374.

Story, J. M., N. W. Callan, J. G. Corn and L. J. White. (2006). Decline of spotted knapweed density at two sites in western Montana with large populations of the introduced root weevil, *Cyphocleonus achates* (Fahraeus). Biological Control.

Sturdevant, N. J., and J. E. Dewey. 2002. Evaluating releases of *Cyphocleonus achates* and *Agapeta zoegana* as potential field insectaries and effects of wildfire on previous releases. USDA Forest Service, Forest Health Protection report 02-05.

Sturdevant, N. J., and J. E. Dewey. 2003. Impact results of *Cyphocleonus achates* on spotted knapweed over 3 years, and corrals as a method for collecting adults-2004. USDA Forest Service, Forest Health Protection report 03-07.